

# Recreational Postrelease Mortality of Lake Trout in Lakes Superior and Huron 

Shawn P. Sitar , Travis O. Brenden, Ji X. He \& James E. Johnson

To cite this article: Shawn P. Sitar , Travis O. Brenden, Ji X. He \& James E. Johnson (2017) Recreational Postrelease Mortality of Lake Trout in Lakes Superior and Huron, North American Journal of Fisheries Management, 37:4, 789-808, DOI: 10.1080/02755947.2017.1327903

To link to this article: http://dx.doi.org/10.1080/02755947.2017.1327903


Accepted author version posted online: 16
May 2017.
Published online: 16 May 2017.


Submit your article to this journal


Article views: 143

View Crossmark data ©

# Recreational Postrelease Mortality of Lake Trout in Lakes Superior and Huron 

Shawn P. Sitar* ${ }^{*}$<br>Michigan Department of Natural Resources, Marquette Fisheries Research Station, 484 Cherry Creek Road, Marquette, Michigan 49855, USA<br>Travis O. Brenden<br>Quantitative Fisheries Center, Department of Fisheries and Wildlife, Michigan State University, 375 Wilson Road, Room 101, East Lansing, Michigan 48824, USA

Ji X. He and James E. Johnson ${ }^{1}$<br>Michigan Department of Natural Resources, Alpena Fisheries Research Station, 160 East Fletcher Street, Alpena, Michigan 49707, USA


#### Abstract

The effectiveness of fishing regulations that result in the release of some angler-caught fish depends on accurate knowledge of the postrelease mortality of those individuals. In the Laurentian Great Lakes, Lake Trout Salvelinus namaycush are a major component of recreational fisheries; across large regions of the lakes, they are managed with length limit and daily quota regulations assuming a $15 \%$ postrelease mortality rate. Due to concerns regarding the accuracy of that rate, we conducted a tagging study to estimate Lake Trout postrelease mortality in Lakes Superior and Huron, and we examined environmental and fishing factors that influenced the return rates of tagged fish. The basic study design was to compare tag return rates between two groups: (1) a treatment group comprising fish that were caught and released by anglers; and (2) a control group comprising fish that were caught via trap net and released. Tag return rates for the angler-caught group were evaluated in relation to depth of capture, surface temperature at release (ST), fishing method, anatomical hook site, play time, handling time, and barotrauma. Tag return rates for angler-caught fish declined significantly with increasing ST; the other factors' effects on tag return rates were generally small. For Lake Superior, model-averaged (Akaike's information criterion) postrelease mortality estimates incorporating ST were $15.0 \%\left(\mathrm{SE}=\mathbf{5 . 6 \%}\right.$ ) at STs less than $10^{\circ} \mathrm{C}, \mathbf{4 2 . 6 \%}(\mathrm{SE}=\mathbf{3 . 0 \%}$ ) at STs of $10-16^{\circ} \mathrm{C}$, and $43.3 \%$ ( $\mathrm{SE}=\mathbf{3 . 6 \%}$ ) at STs greater than $16^{\circ} \mathrm{C}$. Model-averaged estimates for Lake Huron were $\mathbf{5 2 . 5 \%}(\mathrm{SE}=\mathbf{2 6 . 8 \%})$ at STs less than $10^{\circ} \mathrm{C}, \mathbf{4 5 . 2 \%}\left(\mathrm{SE}=\mathbf{1 4 . 0 \%}\right.$ ) at STs of $\mathbf{1 0}-16^{\circ} \mathrm{C}$, and $\mathbf{7 6 . 4 \%}(\mathrm{SE}=\mathbf{5 . 4 \%}$ ) at STs greater than $16^{\circ} \mathrm{C}$. Based on these findings, alternative fishery management regulations that limit recreational catch-and-release angling of Lake Trout in the Great Lakes may be prudent. Current management policies based on an assumed $\mathbf{1 5 \%}$ postrelease mortality are likely underestimating the total numbers of Lake Trout that are removed by recreational anglers.


Size and bag limits are widely used in the regulation of fisheries (Paukert et al. 2001, 2007; Isermann and Paukert 2010) and often result in catch-and-release fishing and grading that can lead to a significant number of fish releases. An example in the Great Lakes
is for Lake Trout Salvelinus namaycush, which are a major component of the recreational fisheries harvest. Great Lakes recreational anglers typically employ downriggers aboard small boats ( $<10 \mathrm{~m}$ ) to catch Lake Trout because they inhabit deep water over

[^0]large areas away from shore. A downrigger is an apparatus that clips to the fishing line above the lure and submerses it to deep water via a heavy weight attached to a cable on a reel (Dedual 1996). In the Great Lakes, downriggers are generally fished at depths between 25 and 60 m , with the vessel traveling less than 5 $\mathrm{km} / \mathrm{h}$. However, some Great Lakes boat anglers catch Lake Trout by trolling, stationary or drift fishing with a weighted line. There is little information on characteristics of the various fishing methods employed by anglers in the Great Lakes, and each method may have different effects on caught fish. Recreational harvest of Lake Trout in the Great Lakes is managed with length limits (Caroffino 2013) and daily quota regulations that have resulted in catch-andrelease angling in some areas (Lockwood et al. 2001; Krueger et al. 2013). In Michigan waters of the upper Great Lakes between 2010 and 2015, total recreational fishery releases were 9,800 fish ( $7 \%$ of catch) in Lake Superior, 16,000 fish ( $18 \%$ of catch) in Lake Huron, and 96,000 fish ( $42 \%$ of catch) in Lake Michigan (T. Kolb, Michigan Department of Natural Resources [MDNR], personal communication). The MDNR angler survey program measures releases of both legal-sized and non-legal-sized Lake Trout, and most releases in Lakes Huron and Michigan during 2010-2015 were related to restrictive length limit regulations, whereas releases in Lake Superior were mostly due to high grading of catch (returning smaller fish when larger fish are caught) because length limits were unrestrictive (MDNR, unpublished data).

Management of Lake Trout is a major focus of Great Lakes natural resources agencies; in many areas, management is supported by routine stock assessments using statistical catch-at-age models that employ fishery harvest and fisheryindependent survey data to estimate population abundances, recruitment levels, and mortality rates. These estimates in turn are used to determine annual harvest quotas based on agreedupon harvest policies (Brenden et al. 2013). A key requirement of statistical catch-at-age analysis is an accurate estimate of total fishery kill, including both actual harvest and the fish that die after release (Quinn and Deriso 1999).

Numerous studies have indicated greater postrelease mortality from catch-and-release fishing practices during high water temperatures (Muoneke and Childress 1994; Bartholomew and Bohnsack 2005; Arlinghaus et al. 2007). Given that the Lake Trout is a coldwater, deepwater species, a similar linkage between postrelease mortality rate and temperature would be expected. Indeed, studies of inland lake recreational fisheries point to temperature as a major determinant of postrelease mortality rates. For inland ice fisheries, estimates of postrelease mortality in Lake Trout have ranged from $9 \%$ to $32 \%$ (Dextrase and Ball 1991; Persons and Hirsch 1994). Similarly, in a Colorado reservoir, estimated Lake Trout postrelease mortality was $12 \%$ during cold temperatures, whereas it was as high as $87 \%$ during the late summer (Lee and Bergersen 1996). In Great Slave Lake, a large oligotrophic lake in northern Canada, $7 \%$ postrelease mortality was estimated for Lake Trout during the open-water fishery when the surface water
temperature was $9^{\circ} \mathrm{C}$ or cooler (Falk et al. 1974). Studies have also pointed to hooking location as an important determinant of resulting postrelease mortality rates, with Lake Trout hooked in vital areas (e.g., gills or stomach) having greater mortality rates than fish hooked in the mouth (Dextrase and Ball 1991; Persons and Hirsch 1994).

Loftus et al. (1988) provided the only estimate of Lake Trout postrelease mortality in the Great Lakes. In that study, charter boat operators and sport boat anglers in Lakes Superior, Huron, and Michigan were employed to catch Lake Trout, and captured fish were tethered for up to 48 h to an anchor-buoy rig. The average postrelease mortality rate from the Loftus et al. (1988) study was $14.9 \%$ ( $95 \%$ confidence interval $=7.4-25.7 \%$ ), although higher levels of postrelease mortality were reported for smaller fish and for fish hooked in vital areas. No effect of depth, temperature differential between the surface and capture depth, lure type, or play time was found in that study.

Based on the results of Loftus et al. (1988), a $15 \%$ postrelease mortality rate has been assumed in harvest policies and regulations enacted for Lake Trout across large areas of Lakes Superior, Huron, and Michigan (Modeling Subcommittee, Technical Fisheries Committee 2002). Nevertheless, concerns about the accuracy of the $15 \%$ estimate have lingered due to perceived limitations in the design of the Loftus et al. (1988) study, including small sample sizes ( 22 fish in year 1; 45 fish in year 2), the limited depth range from which fish were caught ( $<50 \mathrm{~m}$ ), and the short evaluation period (Modeling Subcommittee, Technical Fisheries Committee 2002). Furthermore, barotrauma has been a concern, as most Lake Trout are brought up from deep water and many are observed with overinflated gas bladders (Loftus et al. 1988; Ng et al. 2015).

The objective of our study was to conduct a tagging experiment that allowed us to estimate the postrelease mortality of Lake Trout from the upper Great Lakes and to evaluate how the return rates of tagged fish were affected by factors such as fish length, handling time, play time, surface temperature at the time of release, fishing method, occurrence of barotrauma, and depth of capture. The study was conducted in Lakes Huron and Superior, and we assumed that the results from Lake Huron would be applicable to Lake Michigan because of similarity in limnology (Moll et al. 2013) and angling practices. Although four morphotypes of Lake Trout are extant in Lake Superior (Muir et al. 2014), only the lean morphotype is present in all of the Great Lakes, and it is the form generally targeted by fisheries. All Lake Trout collected in this study belonged to the lean morphotype.

## METHODS

Lake Trout tagging.-For our research, postrelease mortality was evaluated by tagging two groups of Lake Trout: treatment fish (i.e., recreationally angled) and control fish (Pollock and Pine 2007). The treatment group comprised

Lake Trout caught by volunteer boat anglers. In Lake Superior, volunteer boat anglers employed four fishing methods: bobbing (BOB), downrigger with no release (DRNR ), downrigger with release (DR-REL), and wire lining (WIRE; Table 1). In Lake Huron, volunteer boat anglers used three methods: surface fishing (SURF), DR-REL, and WIRE/lead-core (LC) fishing. The control group comprised Lake Trout that were caught in Great Lakes trap nets (Westerman 1932: Brown et al. 1999; Brenden et al. 2013). Trap nets were selected for the control group because earlier research indicated minimal trauma and high survival rates after release from this gear type (Johnson et al. 2004b). Tagging was conducted off two recreational fishing ports of Michigan: Marquette on southern Lake Superior and Alpena on western Lake Huron (Figure 1). These two ports were chosen because of their proximity to research facilities, high levels of recreational harvest and effort for Lake Trout,


FIGURE 1. Study areas (shaded ellipses) where Lake Trout were tagged to assess postrelease mortality in Lakes Superior and Huron.
availability of volunteer anglers, availability of commercial trap-net operators, and high tag return rates as indicated by prior studies. Tagging area boundaries were designated based on prior knowledge of Lake Trout movement and home range patterns (Schmalz et al. 2002; Kapuscinski et al. 2005; Adlerstein et al. 2007). Tagging of both treatment and control groups was restricted to each of the two study areas (Figure 1). Lake Trout were tagged throughout the fishing season (April-November) in 2010-2013. The target annual sample size was 600 fish per study group in each lake but was not achieved in some locations and years. Fish were tagged using serialized, lock-on loop tags (Floy FD-4; Floy Tag and Manufacturing, Inc., Seattle). Except for the unique identification numbers, tags were identical. A US\$10 reward was offered to encourage tag returns. Tags were returned from the recreational fishery, commercial trap-net fishery, commercial gill-net fishery, and natural resource agency gillnet surveys. Tag returns summarized in this paper were collected through June 15, 2016; data used for postrelease mortality estimation were those collected through the end of 2015.

Volunteer boat anglers were recruited at both study areas and were trained on tagging technique, assessment of fish condition, and study protocols for the treatment group. Data collection and tagging techniques were developed such that the treatment fish closely represented actual recreational catch-and-release practices. Data collected for treatment group fish included tag serial number, TL $( \pm 50 \mathrm{~mm})$, date, location, depth of capture (m), play time, handling time, bloating (gas bladder inflated), presence of gulls Larus spp. at the release site, hook location, fishing method, and surface temperature (ST) on the day of tagging. The categorical data collected are described in Table 1. We assessed only the overt symptom of barotrauma by counting fish that were bloated when released and did not document the cryptic symptoms of barotrauma (Wilde 2009). To minimize handling time, digital cameras were used to record much of the data for postprocessing, and electronic chess game timers (Saitek Competition Game Clock, Saitek Industries) were used to record play time and handling time (separately). Each captured fish was placed in a specialized measuring board that restrained the fish and displayed the tag serial number, and a digital photo was taken by the volunteer angler (i.e., the photo recorded the date, tag serial number, and TL). The measuring board comprised a longitudinally sectioned, $152-\mathrm{mm}$-diameter polyvinyl chloride (PVC) pipe that was painted with alternating black-andwhite $50-\mathrm{mm}$ bands so that length group could be measured from the photo. After the fish was tagged and released, a digital photo was also taken of the chess timer, which displayed both the play time and handling time. Hourly ST data were obtained from the online Great Lakes Coastal Forecasting System of the Great Lakes Observing System (2014). Daily mean ST for each tagged and released

TABLE 1. Categorical factors and levels recorded for Lake Trout that were tagged and released by anglers to assess postrelease mortality in Lakes Superior and Huron.

| Factor | Levels |  |
| :--- | :--- | :--- |
| Bloating | Yes or no | Barotrauma indicated by overinflated gas bladder |
| Gulls present | Yes or no | Gulls present in area when tagged fish was released |
| Hook location | Jaw/mouth | Hook embedded in the jaw or outer mouth region |
|  | Eye | Hook embedded in the eye |
|  | Stomach | Hook embedded in the esophagus to stomach region |
|  | Gills | Hook embedded in the gills or gill rakers |
|  | Throat | Hook embedded in the posterior region of the mouth |
|  | Other | Hook embedded in other parts of the body |
| Fishing method | Bobbing (BOB) | Stationary or drift fishing with lure attached to handline or fishing pole (Lake |
|  |  | Superior only) |
|  | Downrigger with no | Lure on leader directly attached to a downrigger cable; vessel trolling |
|  | release (DR-NR) |  |
|  | Downrigger with release | Lure fished from a fishing pole and attached to a downrigger cable with a release |
|  | (DR-REL) | mechanism; vessel trolling |
|  | Wire line/lead core | Lure fished from a fishing pole with a heavy weight and wire line or lead core line; |
|  | (WIRE/LC) | vessel trolling (lead core in Lake Huron only) |
|  | Surface (SURF) | Lure fished from a fishing pole between surface and shallow depths with planer |
|  |  | boards or dipsy divers and no weight; vessel trolling (Lake Huron only) |

Lake Trout was calculated by averaging hourly STs between 0700 and 1600 hours (typical fishing times).

Great Lakes commercial trap nets fished by local commercial operators were used to collect and tag the control group of Lake Trout in the study areas (Figure 1). Tagging was performed by MDNR personnel. Data recorded for control fish included tag serial number, TL (mm), date, location, and depth of capture (m). Any fish collected in the trap net that was not in healthy condition (e.g., bloated) was not tagged and was omitted from the control group. Handling time for trap-net tagged fish was less than 1 min .

Background handling mortality associated with the tagging process was evaluated using hatchery Lake Trout broodstock from the Marquette State Fish Hatchery (Marquette, Michigan). Hatchery Lake Trout were tagged via the same procedures used for both the angler-caught group and the control group. Evaluations of handling mortality were conducted on three groups of fish. The first group, which comprised 20 hatchery Lake Trout selected to be greater than 500 mm TL, was tagged in a training session for volunteer boat anglers during spring 2010. The second group ( $n=60$ fish; mean $\mathrm{TL}=359 \mathrm{~mm}$; range $=251-436 \mathrm{~mm}$ ) and third group ( $n=60$ fish; mean $\mathrm{TL}=739 \mathrm{~mm}$; range $=$ $642-841 \mathrm{~mm}$ ) were tagged by MDNR staff at the hatchery in January 2015. There were no mortalities among group 1 fish at 12 months, and a single mortality (1.7\%) was observed in each of groups 2 and 3 at 6 months. Accordingly, we assumed that handling mortality was minimal and equivalent between the angler-caught group and the
control group. Across groups 1-3, mean handling time was $52 \mathrm{~s}($ range $=27-114 \mathrm{~s})$.

Statistical analysis of factors influencing tag return rates.Individual treatment factors or a combination of factors were evaluated by comparing tag return rates for the angler-caught group with handling time, fishing method, play time, depth of capture, and barotrauma. The statistical tests and post hoc comparisons used for these analyses are described in Table 2. Statistical significance was established at $\alpha=0.05$. Our preliminary analyses detected no differences in tag return rates according to Lake Trout TL due to low sample sizes for small fish ( $<450 \mathrm{~mm}$ ) and large fish ( $>700 \mathrm{~mm}$ ). Therefore, we did not incorporate length in our analyses because of the limited length range of tagged fish. Because we were unable to obtain measurements of temperature at the depth of capture (i.e., to estimate the temperature differential experienced by recreationally caught Lake Trout), we compared tag return rates between ST and depth of capture to gain insight into this effect. We assumed that the temperature differential was low for fish caught in shallower depths and would be greater for fish caught in deeper waters when the lakes were not isothermal. We evaluated simple linear relationships of tag return as a function of ST by $20-\mathrm{m}$ capture depth intervals. A significant negative slope for the greater depth intervals would suggest a potential temperature differential effect.

Estimation of postrelease mortality.-Postrelease mortality for the factors identified as potentially important was

TABLE 2. Statistical tests and post hoc comparisons used to compare tag return rates and depth of capture for Lake Trout in Lakes Huron and Superior.

| Dependent <br> variable | Factor (effect) | Levels |  |
| :--- | :--- | :---: | :--- |
| Tag return rate | Barotrauma <br> Barotrauma, gulls <br> present <br> Fishing method | 2 | Statistical/post hoc test used |

estimated by fitting a multigroup Brownie model (Brownie et al. 1985) to the tag returns of treatment and control fish. More specifically, we used the Hoenig et al. (1998) instantaneous formulation of a Brownie model, as this parameterization was necessary to account for different survival rates among treatment and control fish as a consequence of when tagging was completed during tagging years and the size differences between treatment and control fish. Models were fitted separately for Lakes Huron and Superior. For Lake Superior, two separate Brownie models were fitted to different length-groups of fish (see below).

Following Hoenig et al. (1998), the probability that a tag from the treatment group of Lake Trout would be returned was specified as

where $y=$ year of tagging; $i=$ sampling gear in which a returned fish was caught; $r=\operatorname{tag}$ return year; $\theta=$ postrelease mortality rate; $s_{i}=$ selectivity of the $i$ th fishing gear for treatment fish relative to control fish; $q_{i, r}=$ catchability coefficient for the $i$ th fishing gear in the $r$ th return year; $E_{i, r}=$ amount of effort of the $i$ th fishing gear in the $r$ th return year; $M_{r}=$ instantaneous natural mortality in the $r$ th return year; and $\Delta_{r}$ $=$ length of a period (expressed as a fraction of the year) of the $r$ th return year during which tagged fish were at large in the system. The $\Delta_{r}$ when the return year equaled the year of
tagging was necessary because tagging operations frequently were not completed until sometime during the summer, meaning that the amount of natural mortality experienced by recently tagged fish in that year differed from the natural mortality experienced by previously tagged fish. Similarly, the amount of fishing effort that was specified when the return year equaled the year of tagging was different than for other years to account for the fact that tagging operations were not completed until the summer. The effort measures were anglerhours for the recreational fishery, meters of gill net for the commercial fishery and agency surveys, and number of lifts for commercial trap nets.

In Lake Superior, trap nets tended to catch larger Lake Trout than the volunteer boat anglers. Therefore, we assumed a relative selectivity of 1.25 for the treatment group relative to the control group for returns from recreational angling. Conversely, we assumed a relative selectivity of 0.67 for the treatment group relative to the control group for returns from trap-net gear. For all other fishing gear, equal selectivity was assumed for treatment and control groups. Because there was uncertainty with regard to the selectivities assumed for recreational angling and trap-net gear, we fitted a separate Brownie model to return data for fish that were between 550 and 700 mm TL at tagging-the length range of greatest overlap between the treatment and control groups-to determine the sensitivity of postrelease mortality estimates to differences in gear selectivity. When fitting the Brownie model to fish between 550 and 700 mm TL at the time of tagging, we assumed that selectivity was equal between the sampling gears for fish in the treatment and control groups. For tag returns from Lake Huron, equal selectivity was assumed for all sampling gears because the sizes of Lake

Trout caught by recreational angling and trap nets were similar.

The probability that a tag from a Lake Trout belonging to the control group would be returned was specified using the same equation as used for the treatment group except that (1) postrelease mortality was not included in the equation and (2) the selectivity was set equal to 1.0 for all fishing gears. An additional difference for the control group of fish (in both lakes) was that the return probability was multiplied by 0.984 to account for postrelease mortality based on the results of Johnson et al. (2004a). Reporting rates, handling mortality, and tag retention rates were assumed to be the same for the treatment and control groups. These rates were not factored into the return probabilities, which would lead to biased estimates of natural mortality and catchability from the tagging models but would not influence the estimate of postrelease mortality under the assumption that these rates were the same for the treatment and control groups.

We implemented the tag return models in AD Model Builder (Fournier et al. 2012). Tag returns from both treatment and control groups were modeled through a
multinomial likelihood. Gear catchability and natural mortality rates were estimated on a logarithmic scale to constrain the estimates to positive values. Postrelease mortality rates were estimated through inverse logit functions, which constrained rates between 0.0 and 1.0 while allowing the estimated parameter to occur on the real number line. Diffuse upper and lower bounds were specified for all parameters to prevent the optimization algorithm from flat parts of the objective function surface. Models were considered to have converged on a solution when the maximum gradient of the parameters with respect to the objective function was less than $1.0 \times 10^{-4}$.

We used an information-theoretic approach for evaluating candidate models, which consisted of different combinations wherein postrelease mortality, catchability, and natural mortality varied among the different levels for the factors identified as being important. Evaluated candidate models also included the potential for natural mortality rates from 2010 to 2013 to vary annually (natural mortality rates in 2014 and 2015 were assumed equal to the rate from 2013) and for catchability in the year of tagging to be different from catchability values in other years to account for potential

TABLE 3. Number of Lake Trout tagged, number of tagged fish that were returned, and the tag return rate for the treatment group (caught and released by volunteer recreational anglers) and control group (captured by trap-netting and then released) in each tagging year and return year. Fish in Lake Superior were tagged near the port of Marquette, Michigan, with returns obtained from throughout the lake; fish in Lake Huron were tagged near the port of Alpena, Michigan, and returns were from throughout the lake. Results are based on tag returns recorded through June 15, 2016.

|  | Return year |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Group | Tagging year | Number tagged | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | Number returned | Return rate |
| Lake Superior |  |  |  |  |  |  |  |  |  |  |  |
| Treatment | 2010 | 535 | 32 | 7650 | 32 | 27 | 13 | 8 |  | 188 | 0.351 |
|  | 2011 | 595 |  |  | 76 | 61 | 32 | 9 | 2 | 230 | 0.387 |
|  | 2012 | 590 |  |  | 52 | 67 | 55 | 24 | 7 | 205 | 0.347 |
|  | 2013 | 609 |  |  |  | 29 | 64 | 35 | 11 | 139 | 0.228 |
|  | Total | 2,329 |  |  |  |  |  |  |  | 762 | 0.327 |
| Control | 2010 | 601 | 90 | 100 | 66 | 36 | 11 | 7 |  | 310 | 0.516 |
|  | 2011 | 38 |  | 7 | 6 | 5 |  | 1 |  | 19 | 0.500 |
|  | 2012 | 576 |  |  | 110 | 129 | 51 | 28 | 1 | 319 | 0.554 |
|  | 2013 | 603 |  |  |  | 171 | 99 | 60 | 3 | 333 | 0.552 |
|  | Total | 1,818 |  |  |  |  |  |  |  | 981 | 0.540 |
|  |  |  |  |  | ke Hur | ron |  |  |  |  |  |
| Treatment | 2010 | 249 | 8 | 4 | 3 | 1 | 2 |  |  | 18 | 0.072 |
|  | 2011 | 124 |  | 1 | 2 | 3 | 5 |  |  | 11 | 0.089 |
|  | 2012 | 326 |  |  | 6 | 7 | 2 | 2 |  | 17 | 0.052 |
|  | 2013 | 235 |  |  |  | 1 | 2 | 1 | 1 | 5 | 0.021 |
|  | Total | 934 |  |  |  |  |  |  |  | 51 | 0.055 |
| Control | 2010 | 585 | 60 | 36 | 25 | 6 | 5 | 4 |  | 136 | 0.232 |
|  | 2011 | 459 |  | 40 | 37 | 8 | 4 |  |  | 89 | 0.194 |
|  | 2012 | 310 |  |  | 26 | 20 | 7 | 4 |  | 57 | 0.184 |
|  | 2013 | 317 |  |  |  | 16 | 6 | 1 | 1 | 24 | 0.076 |
|  | Total | 1,671 |  |  |  |  |  |  |  | 306 | 0.183 |




FIGURE 2. Tag return rates by fishing method (defined in Table 1) for Lake Trout tagged by volunteer recreational anglers in Lakes Superior and Huron (bold value within each column = number of tagged individuals from each fishing method). None of the return rates was statistically different within each lake (Marascuilo procedure: $P<0.05$; Appendix Tables A.3, A.4).
nonmixing of tagged fish with at-large populations. Candidate models were evaluated by using Akaike's information criterion (AIC; Burnham and Anderson 2002). For each data set, there was more than one model with AIC difference ( $\triangle \mathrm{AIC}$ ) values less than 10 . To account for model selection uncertainty, model-averaged postrelease mortality estimates and their SEs were calculated from equations given by Burnham and Anderson (2002) based on estimates and AIC weights for all models with $\triangle$ AIC values less than 10 .

Based on analysis of key factors influencing tag return rates, grouped tag return models were fitted, incorporating ST at the time of release as an evaluated factor (see below). We divided ST into three levels $\left(<10,10-16\right.$, and $>16^{\circ} \mathrm{C}$ ) based on results from archival thermal tag studies (Bergstedt et al. 2003, 2016; Mattes 2004; R. Goetz, National Oceanic and Atmospheric Administration [NOAA], Seattle, Washington,


FIGURE 3. Tag return rate as a function of play time interval for Lake Trout that were caught and tagged by volunteer recreational anglers in Lakes Superior and Huron (asterisk $=$ statistically different $[P<0.05]$ tag return rate relative to other play time intervals based on the Marascuilo procedure).
personal communication). Candidate models allowed for postrelease mortality, catchability (potentially time varying), and natural mortality (potentially time varying) to be (1) unique for each ST level, (2) unique for the $<10^{\circ} \mathrm{C}$ ST level but shared between the $10-16^{\circ} \mathrm{C}$ and $>16^{\circ} \mathrm{C}$ ST levels, or (3) shared across all ST levels. In total, 108 models consisting of different combinations of parameters were fitted to the tag return data for each lake.

## RESULTS

## Mark-Recapture of Lake Trout

Between 2010 and 2013, 2,329 Lake Trout in the treatment (angler-caught) group and 1,818 in the control group were tagged in Lake Superior. In Lake Huron, 934 Lake


FIGURE 4. Box plots of capture depth for Lake Trout that were tagged by volunteer recreational anglers in Lakes Superior and Huron compared among different fishing methods (defined in Table 1). The horizontal line in each box indicates the median, the box dimensions represent the interquartile range ( 25 th to 75 th percentiles), the whiskers represent the highest and lowest values within $1.5 \times$ the interquartile range, and the open circles are outliers. Mean depth is indicated by solid triangles. Different letters indicate a statistical difference $(P<0.05)$ between fishing methods based on a Kruskal-Wallis test followed by multiple comparisons using a Nemenyi post hoc test.

Trout in the treatment group and 1,671 fish in the control group were tagged (Table 3). In Lake Superior, there were 10 volunteer boat anglers in 2010 and four volunteer boat anglers in 2011, 2012, and 2013. In Lake Huron, there were
nine volunteer boat anglers during 2010-2012 and seven volunteer boat anglers in 2013. Very few control Lake Trout were tagged during 2011 in Lake Superior because the commercial trap-net operator was unavailable. Overall tag return rates in Lake Superior averaged 54.0\% (range = $50-55.2 \%$ ) for the control group and $32.7 \%$ (range $=22.8-$ $38.7 \%$ ) for the treatment group (Table 3). For Lake Huron, tag return rates averaged $18.3 \%$ (range $=7.6-23.2 \%$ ) for the control group and $5.5 \%$ (range $=2.1-8.9 \%$ ) for the treatment group. Approximately $4 \%$ of tags that were returned had unreadable serial numbers due to tag abrasion and were excluded from analyses.

## Factors Influencing Tag Return Rates

Angler handling times.-Handling time for the majority ( $>65 \%$ ) of Lake Trout tagged by anglers was less than 1.5 min in both Lake Superior and Lake Huron. We compared tag return rates for each fishing method according to five handling time categories ( $<1.0,1.0-1.5,1.5-2.0,2.0-2.5$, and $>2.5 \mathrm{~min}$ ) and found no significant difference in tag return rates for either Lake Superior or Lake Huron (Marascuilo procedure: $P>$ 0.05; Appendix Tables A.1, A.2).

Fishing methods.-For Lake Superior, tag return rates did not differ between fishing methods (Marascuilo procedure: $P$ $>0.05$; Appendix Table A.3; Figure 2). Likewise, in Lake Huron, tag returns were not significantly different between fishing methods (Marascuilo procedure: $P>0.05$; Appendix Table A.4).

Play time.-In Lake Superior, play time for most fish caught by the BOB, DR-NR, and DR-REL methods was 4 min or less. Play time for the WIRE method was more variable, with more than $50 \%$ of fish taking more than 5 min to land. In Lake Huron, play time was no more than 4 min for the majority of fish caught by all fishing methods. We compared tag return rates among six play time intervals ( $<1$, $1-2,2-3,3-4,4-5$, and $>5 \mathrm{~min}$ ) and did not detect significant differences in tag return rate among intervals for any of the fishing methods used in Lake Superior (Marascuilo procedure: $P>0.05$; Appendix Table A.5; Figure 3) except for fish that were caught in less than 1 min by the DR-REL method, which had a significantly lower tag return rate than all other play time intervals (Table A.5; Figure 3). There were no statistical differences in tag return rates according to play time for any of the fishing methods used in Lake Huron (Marascuilo procedure: $P>0.05$; Appendix Table A.6; Figure 3).

Depth of capture.-Overall mean capture depth of tagged Lake Trout belonging to the angler-caught group in Lake Superior was approximately 59 m . Among all angler-caught Lake Superior Lake Trout, the shallowest capture depth was 1.5 m fished by WIRE and the maximum capture depth was 82.3 m fished by DR-REL (Figure 4). In Lake Superior, depth distributions were significantly different between fishing methods (Kruskal-Wallis test: $\chi^{2}=1,240, \mathrm{df}=3, P<$ 0.0001 ). Average capture depth of fish caught by DR-NR


FIGURE 5. Tag return rate by capture depth (m) and fishing method (defined in Table 1) for Lake Trout that were caught and tagged by recreational anglers in Lakes Superior and Huron ( $\mathrm{X}=$ no data for that depth interval).

TABLE 4. Number of Lake Trout tagged, number of tagged fish that were returned, and the tag return rate by depth of capture for fish that were tagged and released by volunteer recreational anglers in Lakes Superior and Huron. Results are based on tag returns recorded through June 15, 2016.

|  | Lake Superior |  |  |  | Lake Huron |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Depth (m) | Number tagged | Number returned | Return rate |  | Number tagged | Number returned | Return rate |
| $<10$ | 4 | 0 | 0.000 |  | 27 | 2 | 0.074 |
| $10-20$ | 22 | 7 | 0.318 |  | 162 | 17 | 0.105 |
| $20-30$ | 11 | 3 | 0.273 |  | 375 | 14 | 0.037 |
| $30-40$ | 42 | 13 | 0.310 |  | 260 | 12 | 0.046 |
| $40-50$ | 513 | 144 | 0.281 |  | 100 | 5 | 0.050 |
| $50-60$ | 863 | 305 | 0.353 | 8 | 1 | 0.125 |  |
| $60-70$ | 265 | 86 | 0.325 | 1 | 0 | 0.000 |  |
| $70-80$ | 224 | 82 | 0.366 |  |  |  |  |
| $>80$ | 368 | 117 | 0.318 |  |  |  |  |

was 47.6 m , and this was the shallowest fishing method (Nemenyi post hoc comparisons: DR-NR versus BOB, $\chi^{2}=$ 312.4, $P<0.0001$; DR-NR versus DR-REL, $\chi^{2}=16.8, P=$ 0.0008 ; DR-NR versus WIRE, $\chi^{2}=13.0, P=0.005$; Figure 4). Mean depths for DR-REL ( 52.3 m ) and $\operatorname{WIRE}(51.7 \mathrm{~m})$ were intermediate among fishing methods and did not statistically differ (Nemenyi post hoc test: $\chi^{2}=0.28, P=0.96$ ). The deepest method of fishing was BOB, with an average capture depth of 78.6 m (Nemenyi post hoc test: $P<0.001$ for all comparisons). For Lake Superior, there was no significant relationship between tag return rate and depth of capture for any of the fishing methods (Figure 5).

In Lake Huron, depth of capture ranged from less than 1 m (surface) for WIRE to 61.6 m for DR-REL. In Lake Huron, the overall mean depth of Lake Trout captured among all fishing methods was 27.3 m . Mean capture depth for the DR-REL method was 28.8 m and was different than that of both the SURF method (Kruskal-Wallis test: $\chi^{2}=$ $144, \mathrm{df}=2, P<0.0001$; Nemenyi post hoc test: $\chi^{2}=46.5, P$ $<0.001$ ) and the WIRE/LC method (Nemenyi post hoc test: $\chi^{2}=107.7, P<0.001$; Figure 4). Mean depth of Lake Trout caught by WIRE/LC was 16.6 m , and mean depth of fish caught by SURF was 20.6 m ; these values did not differ statistically (Nemenyi post hoc test: $\chi^{2}=4.18, P=0.12$ ). There was no significant relationship between tag return rate and depth of capture for any of the fishing methods in Lake Huron (Figure 5). In both Lake Superior and Lake Huron, there were no statistical differences in tag return rate relative to depth of capture for all fishing methods combined (Marascuilo procedure: $P>0.05$; Table 4; Appendix Tables A.7, A.8).

Barotrauma.-Bloating of angler-caught Lake Trout was observed in $32.3 \%$ of Lake Superior fish and only $5.6 \%$ of fish in Lake Huron. Incidence of barotrauma was related to the depth of capture: bloating was significantly more frequent for Lake Trout caught at $50-\mathrm{m}$ or greater depths in Lake Superior ( $Z$-test: $Z=-3.15, P=0.002$ ) and for fish caught at


FIGURE 6. Influence of barotrauma (indicated by bloating) and the presence of gulls on the tag return rate for Lake Trout that were tagged by volunteer recreational anglers in Lakes Superior and Huron (bold value within each column = number of tagged individuals). White columns represent return rates for fish that were released without gulls present; gray columns represent return rates for fish that were released in the presence of gulls (asterisk $=$ statistically different tag return rate for gull presence; $2 \times 2$ contingency table analysis: $P<0.05$ ).

POSTRELEASE MORTALITY OF LAKE TROUT

TABLE 5. Number of Lake Trout tagged, number of tagged fish that were returned, and the tag return rate by anatomical hooking location (defined in Table 1) for fish that were tagged and released by volunteer recreational anglers in Lakes Superior and Huron. Results are based on tag returns recorded through June 15, 2016.

|  | Anatomical hooking location |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Statistic | Eye | Gills | Jaw | Stomach | Throat | Other |
|  |  | Lake Superior |  |  |  |  |
| Number tagged | 43 | 25 | 2,180 | 2 | 2 | 59 |
| Number returned | 10 | 3 | 716 | 1 | 1 | 24 |
| Tag return rate | 0.23 | 0.12 | 0.33 | 0.50 | 0.50 | 0.41 |
|  |  |  | Lake Huron |  |  |  |
| Number tagged | 1 | 3 | 923 | 1 | 2 | 0 |
| Number returned | 0 | 0 | 51 | 0 | 0 | 0 |
| Tag return rate | 0.00 | 0.00 | 0.06 | 0.00 | 0.00 | 0.00 |

$40-\mathrm{m}$ or greater depths in Lake Huron $(Z=-4.83, P<$ 0.001 ). Gulls were present at the time of release for $4.8 \%$ of Lake Trout in Lake Superior and $2.9 \%$ of fish in Lake Huron. Overall tag return rates for bloated fish did not differ from those of non-bloated fish in both Lake Superior (Z-test: $Z=1.33, P=0.184)$ and Lake Huron $(Z=0.541, P=0.59$; Figure 6). In Lake Superior, tagged Lake Trout that were bloated and that had gulls present at the time of release exhibited a significantly a lower tag return rate than bloated fish with no gulls present or non-bloated fish $(2 \times 2$ contingency table: $P \leq 0.05$; Figure 6). For Lake Huron, no statistical differences in tag return rates based on barotrauma or gull presence/absence were detected.

Anatomical hook location.-Most of the angler-tagged fish ( $94.3 \%$ in Lake Superior; $98.9 \%$ in Lake Huron) were hooked in the jaw or mouth (Table 5). Fish in the "other" category were reported to be hooked on the non-vital parts of the outer body, such as the tail, head, fins, and musculature, and had a tag return rate that was not significantly different than that of fish hooked in the jaw or mouth ( $Z$-test: $Z=-1.29, P=0.197$ ). For all fishing methods combined in Lake Superior, the tag return rate for fish hooked in the eyes or gills (pooled data) was significantly lower than that of fish hooked in the jaw or mouth ( $Z=2.43, P=0.015$ ). In Lake Huron, tag return rates were not significantly different between fish hooked in the jaw or mouth and those hooked in other body locations ( $Z=0.799$, $P=0.424$ ).

Surface temperature at release.-Lake Trout in Lake Superior were tagged throughout the fishing season during April-November and were released in STs ranging from $3^{\circ} \mathrm{C}$ to $23^{\circ} \mathrm{C}$ (Figure 7). In Lake Huron, the fishing season spanned April-October, with an ST range of $7-24^{\circ} \mathrm{C}$ (Figure 7). Overall, Lake Trout were released in warmer temperatures in Lake Huron than in Lake Superior. For Lake Superior, the full ANCOVA model evaluating tag return rate as a function of ST and that included year and
group resulted in no significant interactions of $\mathrm{ST} \times$ group $\left(F_{1,54}=1.01, P=0.32\right)$, ST $\times$ year $\left(F_{2,54}=0.23, P=\right.$ $0.795)$, group $\times$ year $\left(F_{2,54}=1.18, P=0.314\right)$, or ST $\times$ year $\times \operatorname{group}\left(F_{2,54}=0.19, P=0.824\right)$. Furthermore, there was no significant year effect $\left(F_{2,54}=2.79, P=0.07\right)$. In the reduced model, no significant interaction of $\mathrm{ST} \times$ group was detected $\left(F_{1,62}=0.79, P=0.379\right)$. For the angler-caught group, tag return rates declined significantly with increasing ST (intercept: $t=9.982, P<0.001$; slope: $t=-3.83, P=$ 0.0003 ; Figure 8). There was no was significant relationship of tag return as a function of ST for the trap-netted group (intercept: $t=0.56, P=0.577$; slope: $t=0.89, P=0.379$; Figure 8). For Lake Huron, no significant relationship between ST and tag return rate was detected $\left(F_{1,75}=1.00\right.$, $P=0.321$; Figure 8).

For Lake Superior, significant negative relationships between tag return rate and ST were found for Lake Trout caught at depths of $40-60 \mathrm{~m}\left(F_{1,18}=5.89, P=0.026\right), 60-80$ $\mathrm{m}\left(F_{1,18}=30.1, P<0.0001\right)$, and over $80 \mathrm{~m}\left(F_{1,19}=31.6, P\right.$ $<0.0001$; Table A.7; Figure 9). In shallower waters, no significant relationship between tag return rate and ST was detected for Lake Trout caught at depths less than $20 \mathrm{~m}\left(F_{1}\right.$, $\left.{ }_{6}=2.26, P=0.183\right)$ and $20-40 \mathrm{~m}\left(F_{1,9}=0.21, P=0.658\right)$. Only $3.5 \%$ of angler-caught fish were captured at depths less than 40 m in Lake Superior. For Lake Huron, no significant relationship was observed between tag return rate and ST according to depth of capture ( $<20 \mathrm{~m}: F_{1,14}=0.006, P=$ $0.937 ; 20-40 \mathrm{~m}: F_{1,16}=0.062, P=0.806 ; 40-60 \mathrm{~m}: F_{1,8}=$ $0.764, P=0.408$; Table A.8).

## Estimation of Postrelease Mortality

For Lake Superior, there were 12 models with $\triangle \mathrm{AIC}$ values less than 10 for both the full data set and the reduced data set (i.e., limited to fish between 550 and 700 mm TL at the time of tagging; Table 6). The models with $\triangle$ AIC values less than 10 were the same for both data sets, although there


FIGURE 7. Box plot of surface temperature $\left(\mathrm{ST} ;{ }^{\circ} \mathrm{C}\right)$ at the time of release for Lake Trout that were tagged by volunteer recreational anglers in Lakes Superior and Huron during 2010-2013. The horizontal line in each box indicates the median, the box dimensions represent the interquartile range (25th to 75th percentiles), the whiskers represent the highest and lowest values within $1.5 \times$ the interquartile range, and the open circles are outliers. Mean monthly ST is represented by solid triangles.
were slight variations in model rankings between the data sets. Six of the 12 best-performing models, including the model with the overall lowest AIC value, for both data sets estimated a unique postrelease mortality rate for the low-ST group ( $<10^{\circ} \mathrm{C}$ ) and a shared postrelease mortality rate for the medium-ST $\left(10-16^{\circ} \mathrm{C}\right)$ and high-ST $\left(>16^{\circ} \mathrm{C}\right)$ groups


FIGURE 8. Relationship between tag return rate $(R)$ and surface temperature ( $\mathrm{ST} ;{ }^{\circ} \mathrm{C}$ ) for angler-caught Lake Trout (tag return rate $R_{A}$ ) and control fish (captured via trap-netting) that were tagged in Lakes Superior and Huron during 2010-2013.
(Table 6). The other six models with $\Delta$ AICs below 10 estimated a unique postrelease mortality for each ST group (Table 3). Across the different models, variation in postrelease mortality estimates was generally small, as the absolute difference in postrelease mortality estimates between models within a particular ST level was no greater than $4.3 \%$ for both data sets (Table 3).

For Lake Huron, there were 34 models with $\triangle \mathrm{AIC}$ values less than 10 (Table 6). The six best-performing models, which all had $\triangle$ AICs less than 4 , estimated a unique postrelease mortality rate for each ST group. Compared to Lake Superior, there was greater variation in postrelease mortality estimates among models within the ST groups for Lake Huron. The largest absolute difference in postrelease mortality estimates between models within the ST groups was $21.8 \%$


FIGURE 9. Relationship between the tag return rate $(R)$ and surface temperature $\left(\mathrm{ST} ;{ }^{\circ} \mathrm{C}\right)$ by depth of capture for Lake Trout that were caught and released by volunteer recreational anglers in Lake Superior $\left(R^{2}=\right.$ coefficient of determination).
$\left(\mathrm{ST}<10^{\circ} \mathrm{C}\right), 30.9 \%\left(\mathrm{ST}=10-16^{\circ} \mathrm{C}\right)$, and $12.9 \%\left(\mathrm{ST}>16^{\circ} \mathrm{C}\right.$; Table 6).

The model-averaged postrelease mortality estimates for the ST groups based on the full Lake Superior data set (with all lengths of fish) were $15.0 \%(\mathrm{SE}=5.6 \%)$ for STs less than $10^{\circ} \mathrm{C}, 42.6 \%(\mathrm{SE}=3.0 \%)$ for STs of $10-16^{\circ} \mathrm{C}$, and $43.3 \%$ $(\mathrm{SE}=3.6 \%)$ for STs greater than $16^{\circ} \mathrm{C}$. For the reduced Lake Superior data set ( $550-700-\mathrm{mm}$ fish), the model-averaged postrelease mortality estimates were $13.7 \%$ ( $\mathrm{SE}=6.6 \%$ ) for STs less than $10^{\circ} \mathrm{C}, 48.5 \%$ ( $\mathrm{SE}=3.4 \%$ ) for STs of $10-16^{\circ} \mathrm{C}$, and $48.4 \% ~(\mathrm{SE}=3.9 \%)$ for STs greater than $16^{\circ} \mathrm{C}$. Modelaveraged postrelease mortality estimates for Lake Huron were $52.4 \%$ ( $\mathrm{SE}=26.8 \%$ ) for STs below $10^{\circ} \mathrm{C}, 45.2 \%$ ( $\mathrm{SE}=$ $14.0 \%$ ) for STs of $10-16^{\circ} \mathrm{C}$, and $76.4 \%$ ( $\mathrm{SE}=5.4 \%$ ) for STs exceeding $16^{\circ} \mathrm{C}$.

## DISCUSSION

The postrelease mortality we estimated for Lake Trout in the Great Lakes was greater than that estimated by Loftus et al. (1988). The key factor influencing postrelease mortality from recreational fishing was high ST at the time of capture. Postrelease mortality estimates were generally consistent between Lakes Superior and Huron for angler-tagged fish released in STs between $10^{\circ} \mathrm{C}$ and $16^{\circ} \mathrm{C}$. For fish released in STs less than $10^{\circ} \mathrm{C}$, the postrelease mortality
estimate from Lake Huron was greater than that from Lake Superior but also had greater uncertainty, which was due at least partly to the low number of recaptures $(n=3$ for angler-caught fish) at this ST level (only 39 tagged fish were released in $\mathrm{STs}<10^{\circ} \mathrm{C}$ among all years). For fish captured and released at high $\mathrm{STs}\left(>16^{\circ} \mathrm{C}\right)$, the greater postrelease mortality in Lake Huron may have been driven by the difference in temperature distributions between lakes. Among Lake Trout that were released in STs over $16^{\circ} \mathrm{C}$, the majority of fish were tagged and released between $17^{\circ} \mathrm{C}$ and $19^{\circ} \mathrm{C}$ in Lake Superior, whereas they were released in STs between $19^{\circ} \mathrm{C}$ and $24^{\circ} \mathrm{C}$ in Lake Huron. In Lake Superior, postrelease mortality rates at $\mathrm{STs} \geq 10^{\circ} \mathrm{C}$ were more than 2.5 times the mortality rates observed at STs less than $10^{\circ} \mathrm{C}$. In Lake Huron, postrelease mortality rates at $\mathrm{STs} \geq 10^{\circ} \mathrm{C}$ were approximately 1.5 times those at STs less than $10^{\circ} \mathrm{C}$.

From laboratory experiments, optimal thermal habitat for Lake Trout has been reported to be between $8^{\circ} \mathrm{C}$ and $12^{\circ} \mathrm{C}$ (Christie and Regier 1988; Magnuson et al. 1990; Mackenzie-Grieve and Post 2006). More recent archival thermal tagging studies for Lake Trout in Lake Huron (Bergstedt et al. 2003, 2016) and in Lake Superior (Mattes 2004; R. Goetz, NOAA, personal communication) indicate that Lake Trout may spend short periods in waters warmer than $10^{\circ} \mathrm{C}$ but spend the bulk of their time in water temperatures less than $10^{\circ} \mathrm{C}$. The causative mechanism for greater postrelease mortality at high STs may be the compound effect of (1) the temperature differential experienced by Lake Trout when brought up from deep, cold waters to warm surface temperatures that are unsuitable for this species; combined with (2) the stress of being hooked, dragged, and reeled in by anglers. Angling is known to induce negative physiological effects in fish by elevating stress hormones and lactate levels (Lee and Bergersen 1996; Morrissey et al. 2005; Tracey et al. 2016). In our study, control Lake Trout (trap-netted group) released in warm temperatures were able to survive better than angler-caught fish because minimal trauma was experienced by trapnetted fish.

An unexpected result in this study was that neither the occurrence of bloating nor the depth of capture had any effect on tag return rates. Capture depth and bloating have been found to affect survival in a variety of species, including the Walleye Sander vitreus (Talmage and Staples 2011), Largemouth Bass Micropterus salmoides (Feathers and Knable 1983), and Striped Bass Morone saxatilis (Bettoli and Osborne 1998). Possible explanations for the lack of observed effect from bloating or depth of capture are that (1) the depth effect was confounded with temperature (as discussed above) and (2) Lake Trout are physostomous, and some bloated fish were able to recover by decompression of the gas bladder, which allowed them to return to deeper waters after release ( Ng et al. 2015). This was observed by Loftus et al. (1988) and by volunteer anglers in this study,

TABLE 6. Akaike's information criterion (AIC) values, number of parameters ( $K$ ), and postrelease mortality estimates (\%) for Lake Trout released at three surface temperature (ST) levels (high [H], $>16^{\circ} \mathrm{C}$; medium [M], $10-16^{\circ} \mathrm{C}$; low [L], $<10^{\circ} \mathrm{C}$ ) based on tag return models fitted to each lake and data set. Only models with AIC difference ( $\triangle$ AIC) values less than 10 are shown. Models are identified based on whether the postrelease mortality rate $(j)$, recreational fishing gear catchability ( $q$ ), and/or natural mortality rate ( $M$ ) differed by ST group or were time varying (TV). For an least some of the ST levels (i.e., the dot replaces the ST levels for which coefficients were shared). For $q$, "TV" indicates a model for which $q$ in the year of tagging differed from that in other tag return years. For $M$, "TV" indicates a model in which $M$ values during 2010-2013 differed annually but $M$ values in 2014 and 2015 were set equal to the 2013 level. Blank cells for Lake Huron indicate that models were not fitted to a reduced data set (i.e., $550-700-\mathrm{mm}$ fish).

| Model label | $K$ | All data |  |  |  | 550-700-mm tagging length |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | AIC | H | M | L | AIC | H | M | L |
| Lake Superior |  |  |  |  |  |  |  |  |  |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\cdot)$ | 27 | 13,997.3 | 42.4 | 42.4 | 14.5 | 10,375.2 | 48.2 | 48.2 | 12.8 |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\cdot)$ | 28 | 13,999.1 | 44.2 | 41.7 | 14.5 | 10,377.2 | 48.2 | 48.2 | 12.8 |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\cdot, \mathrm{~L})$ | 28 | 13,999.3 | 42.4 | 42.4 | 14.5 | 10,376.7 | 48.0 | 48.0 | 13.5 |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot \mathrm{L})$ | 37 | 14,001.0 | 44.3 | 44.3 | 16.5 | 10,379.0 | 49.6 | 49.6 | 16.6 |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\cdot, \mathrm{~L})$ | 29 | 14,001.1 | 44.2 | 41.7 | 14.5 | 10,378.7 | 47.8 | 48.0 | 13.5 |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 29 | 14,001.3 | 42.4 | 42.4 | 14.5 | 10,378.6 | 47.8 | 47.8 | 13.5 |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot)$ | 36 | 14,001.3 | 44.3 | 44.3 | 17.4 | 10,378.9 | 50.4 | 50.4 | 15.5 |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 38 | 14,001.6 | 44.2 | 44.2 | 16.5 | 10,379.5 | 49.5 | 49.5 | 16.6 |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot \mathrm{L})$ | 38 | 14,002.7 | 46.7 | 43.3 | 16.5 | 10,381.0 | 49.9 | 49.5 | 16.6 |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot)$ | 37 | 14,002.9 | 46.5 | 43.3 | 17.5 | 10,380.9 | 50.7 | 50.3 | 15.5 |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 30 | 14,003.1 | 44.2 | 41.7 | 14.5 | 10,380.5 | 46.9 | 48.1 | 13.5 |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 39 | 14,003.6 | 44.4 | 44.2 | 16.5 | 10,381.4 | 48.2 | 50.0 | 16.6 |
| Lake Huron |  |  |  |  |  |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot \mathrm{L})$ | 38 | 3,571.7 | 77.3 | 40.0 | 46.1 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\cdot), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 10 | 3,572.1 | 77.4 | 44.5 | 59.0 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\cdot), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 19 | 3,572.2 | 77.5 | 46.6 | 55.1 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 39 | 3,572.6 | 77.8 | 39.4 | 46.1 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \cdot \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 35 | 3,574.3 | 77.3 | 44.5 | 46.1 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\cdot, \mathrm{~L})$ | 29 | 3,575.5 | 77.3 | 38.4 | 51.7 |  |  |  |  |
| $j(\cdot), q(\cdot), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 17 | 3,576.4 | 67.6 | 67.6 | 67.6 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 30 | 3,576.6 | 77.7 | 37.4 | 51.7 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\cdot), M(\mathrm{TV}, \cdot \mathrm{L})$ | 18 | 3,577.0 | 74.9 | 52.8 | 55.0 |  |  |  |  |
| $j(\cdot), q(\cdot), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 8 | 3,577.0 | 67.4 | 67.4 | 67.4 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\cdot, \mathrm{~L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 23 | 3,577.8 | 77.7 | 47.0 | 51.6 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \cdot \mathrm{L}), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 26 | 3,577.9 | 77.1 | 42.1 | 51.7 |  |  |  |  |
| $j(\cdot, \mathrm{~L}), q(\cdot), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 18 | 3,578.1 | 68.2 | 68.2 | 55.1 |  |  |  |  |
| $j(\cdot), q(\cdot), M(\mathrm{TV}, \cdot \mathrm{L})$ | 16 | 3,578.1 | 66.8 | 66.8 | 66.8 |  |  |  |  |
| $j(\cdot), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot \mathrm{L})$ | 36 | 3,578.3 | 66.1 | 66.1 | 66.1 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\cdot, \mathrm{~L}), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 14 | 3,578.6 | 77.5 | 44.8 | 57.6 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\cdot), M(\mathrm{TV}, \cdot)$ | 17 | 3,578.8 | 75.8 | 53.8 | 52.0 |  |  |  |  |
| $j(\cdot, \mathrm{~L}), q(\cdot), M(\mathrm{H}, \mathrm{M}, \mathrm{L})$ | 9 | 3,578.9 | 67.9 | 67.9 | 59.0 |  |  |  |  |

TABLE 6. Continued.

| Model label | K | All data |  |  |  | 550-700-mm tagging length |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | AIC | H | M | L | AIC | H | M | L |
| $j(\cdot), q(\mathrm{TV}, \cdot, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 33 | 3,579.4 | 66.7 | 66.7 | 66.7 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \cdot, \mathrm{L}), M(\mathrm{TV}, \cdot, \mathrm{L})$ | 34 | 3,579.5 | 74.2 | 51.0 | 46.1 |  |  |  |  |
| $j(\cdot), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 37 | 3,579.7 | 66.1 | 66.1 | 66.1 |  |  |  |  |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot, \mathrm{L})$ | 37 | 3,579.7 | 67.0 | 67.0 | 46.1 |  |  |  |  |
| $j(\cdot, \mathrm{~L}), q(\cdot), M(\mathrm{TV}, \cdot, \mathrm{L})$ | 17 | 3,579.8 | 67.4 | 67.4 | 55.0 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot \mathrm{L})$ | 26 | 3,580.2 | 77.4 | 40.5 | 51.6 |  |  |  |  |
| $j(\cdot), q(\cdot), M(\mathrm{TV}, \cdot)$ | 15 | 3,580.3 | 67.9 | 67.9 | 67.9 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot)$ | 37 | 3,580.5 | 78.1 | 39.4 | 47.7 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 27 | 3,580.5 | 78.2 | 40.1 | 51.6 |  |  |  |  |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \cdot, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 34 | 3,580.8 | 67.7 | 67.7 | 46.1 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \cdot)$ | 28 | 3,581.0 | 78.4 | 39.0 | 51.7 |  |  |  |  |
| $j(\cdot, \mathrm{~L}), q(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L}), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 38 | 3,581.0 | 67.0 | 67.0 | 46.1 |  |  |  |  |
| $j(\cdot), q(\mathrm{TV}, \cdot, \mathrm{L}), M(\mathrm{TV}, \cdot, \mathrm{L})$ | 32 | 3,581.1 | 65.5 | 65.5 | 65.5 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\cdot), M(\mathrm{TV}, \cdot \mathrm{L})$ | 9 | 3,581.3 | 74.1 | 53.6 | 58.8 |  |  |  |  |
| $j(\mathrm{H}, \mathrm{M}, \mathrm{L}), q(\mathrm{TV}, \cdot), M(\mathrm{TV}, \mathrm{H}, \mathrm{M}, \mathrm{L})$ | 31 | 3,581.5 | 75.7 | 43.5 | 52.6 |  |  |  |  |
| $j(\cdot), q(\cdot), M(\mathrm{TV}, \cdot, \mathrm{L})$ | 7 | 3,581.5 | 66.5 | 66.5 | 66.5 |  |  |  |  |

suggesting that there is little benefit of decompressing the gas bladders of bloated Lake Trout because even though they have the ability to recover on their own, the fish are already compromised from the overall trauma associated with recreational catch. The one potential exception is when bloating occurs in the presence of gulls, because there did appear to be some combined effect of bloating and gulls on tag return rates of Lake Trout in Lake Superior, although a similar effect was not observed for fish in Lake Huron.

Prior research has indicated significantly higher postrelease mortality for Lake Trout smaller than 509 mm ( $57.1 \%$; Loftus et al. 1988). Unfortunately, our study lacked a sufficient sample size of smaller Lake Trout to assess this. Since size limit regulations are important tools for managing Lake Trout, an examination of postrelease mortality according to size would be valuable for future research.

Based on the results of this research, the use of recreational catch-and-release practices with Lake Trout in the Great Lakes poses a management dilemma. Most Lake Trout in the upper Great Lakes are harvested during the summer months, when STs are well above the species' thermal optimum. For example, during our study period (2010-2015), $76 \%$ of total recreational harvest in Lake Superior and $97.5 \%$ of recreational harvest in Lake Huron occurred during months when STs were $10^{\circ} \mathrm{C}$ or higher. Regulations that require Great Lakes anglers to release Lake Trout will have a limited protective effect, as $40-76 \%$ of released fish may not survive and perhaps even higher mortality rates will occur during warmer months. Lake Trout that are subject to recreational catch and release are physiologically compromised, and the scope for recovery and survival is limited by release into suboptimal STs. It is apparent that Lake Trout may not be suitable for recreational catch-and-release fishing in the Great Lakes. Restrictive recreational length limits for Lake Trout may not produce the desired management outcome; therefore, it would be prudent for resource agencies to consider alternatives that would minimize overall catch, such as season or area restrictions or limiting daily quotas. Current management policies based on an assumed $15 \%$ postrelease mortality rate are likely underestimating the total numbers of fish harvested by recreational anglers, and we recommend updating the assumed postrelease mortality rates based on the present results.

## ACKNOWLEDGMENTS

We thank the following staff at MDNR research stations for supporting this study: Penny Bacon, Ed Barr, Steve Dewitt, Kevin Duby, Ken Glomski, Andy Jasonowicz, Deb Macconnell, Eric Mammoser, Karen Sanford, Nick Steimel, Dan Traynor, Darren Vercnocke, and Bill Wellenkamp. We greatly appreciate the volunteer anglers
and commercial fishers who made this study possible: for Lake Superior, Joe Buys, Rick Sarasien, Roy Isaacson, Sam and Kathy L'Huillier, Neil Green, Joe December, Joe Gerbyshak, and Thill Fisheries; and for Lake Huron, Brad Valley, Bryan Lapine, Bryan Valley, Chris Klein, Dick Rang, Ed Retherford, Ernest Andree, Jason Snyder, Jason Witkowski, Jerry Perrin, Mike Berend, Scott Gauthier, Stephen Alexander, Steve Speaks, Terry Wortley, Rochefort Fisheries, and Spaulding-Gauthier Fisheries. We thank Mark Ebener for suggesting the study design and providing guidance throughout the project. This study was funded by the MDNR through Federal Aid in Sport Fish Restoration Project F-80-R. Additional funding support was provided by contributing partners of the Quantitative Fisheries Center at Michigan State University.

## ORCID

Shawn P. Sitar (1) http://orcid.org/0000-0002-8868-3277

## REFERENCES

Adlerstein, S. A., E. S. Rutherford, J. A. Clevenger, J. E. Johnson, D. F. Clapp, and A. P. Woldt. 2007. Lake Trout movements in U.S. waters of Lake Huron interpreted from coded wire tag recoveries in recreational fisheries. Journal of Great Lakes Research 33:186-201.
Arlinghaus, R., S. J. Cooke, J. Lyman, D. Policansky, A. Schwab, C. Suski, S. G. Sutton, and E. B. Thorstad. 2007. Understanding the complexity of catch-and-release in recreational fishing: an integrative synthesis of global knowledge from historical, ethical, social, and biological perspectives. Reviews in Fisheries Science 15:75-167.
Bartholomew, A., and J. A. Bohnsack. 2005. A review of catch-and-release angling mortality with implications for no-take reserves. Reviews in Fish Biology and Fisheries 15:129-154.
Bergstedt, R. A., R. L. Argyle, J. G. Seelye, K. T. Scribner, and G. L. Curtis. 2003. In situ determinations of the annual thermal habitat use by Lake Trout (Salvelinus namaycush) in Lake Huron. Journal of Great Lakes Research 29(Supplement 1):347-361.
Bergstedt, R. A., R. L. Argyle, W. W. Taylor, and C. Krueger. 2016. Seasonal and diel bathythermal distributions of Lake Whitefish in Lake Huron: potential implications for Lake Trout bycatch in commercial fisheries. North American Journal of Fisheries Management 36:705-719.
Bettoli, P. W., and R. S. Osborne. 1998. Hooking mortality and behavior of Striped Bass following catch and release angling. North American Journal of Fisheries Management 18:609-615.
Brenden, T. O., R. W. Brown, M. P. Ebener, K. Reid, and T. J. Newcomb. 2013. Great Lakes commercial fisheries: historical overview and prognoses for the future. Pages 339-397 in W. W. Taylor, A. Lynch, and N. Leonard, editors. Great Lakes fisheries policy and management: a binational perspective, 2nd edition. Michigan State University Press, East Lansing.
Brown, R. W., M. Ebener, and T. Gorenflo. 1999. Great Lakes commercial fisheries: historical overview and prognoses for the future. Pages 307-354 in W. W. Taylor and C. P. Ferreri, editors. Great Lakes fisheries policy and management: a binational perspective. Michigan State University Press, East Lansing.
Brownie, C., D. R. Anderson, K. P. Burnham, and D. S. Robson. 1985. Statistical inference from band recovery data: a handbook. U.S. Fish and Wildlife Service, Resource Publication 156, Washington, D.C.
Burnham, K. P., and D. R. Anderson. 2002. Model selection and multimodel inference: a practical information-theoretic approach, 2nd edition. Springer-Verlag, Fort Collins, Colorado.

Burnham, K. P., D. R. Anderson, G. C. White, C. Brownie, and K. H. Pollock. 1987. Design and analysis methods for fish survival experiments based on release-recapture. American Fisheries Society, Monograph 5, Bethesda, Maryland.
Caroffino, D. C. 2013. Angler compliance with Lake Trout length limit regulations in Great Lakes waters. North American Journal of Fisheries Management 33:1203-1209.
Christie, G. C., and H. A. Regier. 1988. Measures of optimal thermal habitat and their relationship to yields for four commercial fish species. Canadian Journal of Fisheries and Aquatic Sciences 45:301-314.
Dedual, M. 1996. Observed mortality of Rainbow Trout caught by different angling techniques in Lake Taupo, New Zealand. North American Journal of Fisheries Management 16:357-363.
Dextrase, A. J., and H. E. Ball. 1991. Hooking mortality of Lake Trout angled through the ice. North American Journal of Fisheries Management 11:477-479.
Falk, M. R., D. V. Gillman, and L. W. Dahlke. 1974. Comparison of mortality between barbed and barbless hooked Lake Trout. Environment Canada, Fisheries and Marine Service, Report CEN/T-74-1, Winnipeg, Manitoba.
Feathers, M. G., and A. E. Knable. 1983. Effects of depressurization upon Largemouth Bass. North American Journal of Fisheries Management 3:86-90.
Fournier, D. A., H. J. Skaug, J. Ancheta, J. Ianelli, A. Magnusson, M. N. Maunder, A. Nielsen, and J. Sibert. 2012. AD Model Builder: using automatic differentiation for statistical inference of highly parameterized complex nonlinear models. Optimization Methods and Software 27:233-249.
Great Lakes Observing System. 2014. Lake Superior and Lake Huron Nowcast 3D from the NOAA Great Lakes Coastal Forecast System [online database]. Great Lakes Observing System, Ann Arbor, Michigan. Available: http://data.glos.us/glcfs. (May 2017).
Hoenig, J. M., N. J. Barrowman, W. S. Hearn, and K. H. Pollock. 1998. Multiyear tagging studies incorporating fishing effort data. Canadian Journal of Fisheries and Aquatic Sciences 55:1466-1476.
Isermann, D. A., and C. P. Paukert. 2010. Regulating harvest. Pages 185-212 in W. A. Hubert and M. C. Quist, editors. Inland fisheries management in North America, 3rdedition. American Fisheries Society, Bethesda, Maryland.
Johnson, J. E., M. P. Ebener, K. Gebhardt, and R. Bergstedt. 2004a. Comparison of catch and Lake Trout bycatch in commercial trap nets and gill nets targeting Lake Whitefish in northern Lake Huron. Michigan Department of Natural Resources, Fisheries Research Report 2071, Lansing.
Johnson, J. E., J. L. Jonas, and J. W. Peck. 2004b. Management of commercial fisheries bycatch, with emphasis on Lake Trout fisheries of the upper Great Lakes. Michigan Department of Natural Resources, Fisheries Research Report 2070, Lansing.
Kapuscinski, K. L., M. J. Hansen, and S. T. Schram. 2005. Movements of Lake Trout in U.S. waters of Lake Superior, 1973-2001. Transactions of the American Fisheries Society 25:696-708.
Krueger, C. C., S. R. LaPan, C. R. Schneider, and T. H. Eckert. 2013. Regulation of sport fishery harvest of Lake Trout: use of size limits in New York's waters of Lake Ontario. Pages 589-607 in W. W. Taylor, A. Lynch, and N. Leonard, editors. Great Lakes fisheries policy and management: a binational perspective, 2nd edition. Michigan State University Press, East Lansing.
Lee, W. C., and E. P. Bergersen. 1996. Influence of thermal and oxygen stratification on Lake Trout hooking mortality. North American Journal of Fisheries Management 16:175-181.
Lockwood, R. N., J. W. Peck, and J. Oelfke. 2001. Survey of angling in Lake Superior waters at Isle Royale National Park, 1998. North American Journal of Fisheries Management 21:471-481.
Loftus, A. J., W. W. Taylor, and M. Keller. 1988. An evaluation of Lake Trout (Salvelinus namaycush) hooking mortality in the upper Great Lakes. Canadian Journal of Fisheries and Aquatic Sciences 45:1473-1479.
Mackenzie-Grieve, J. L., and J. R. Post. 2006. Thermal habitat use by Lake Trout in two contrasting Yukon Territory lakes. Transactions of the American Fisheries Society 135:727-738.

Magnuson, J. J., J. D. Meisner, and D. K. Hill. 1990. Potential changes in the thermal habitat of Great Lakes fish after global climate warming. Transactions of the American Fisheries Society 119:254-264.
Marascuilo, L. A. 1966. Large-scale multiple comparisons. Psychological Bulletin 69:280-290.
Mattes, W. P. 2004. Temperature and depth profiles of namaycush (Lake Trout) in Lake Superior. Great Lakes Indian Fish and Wildlife Commission, Project Report 04-01, Odanah, Wisconsin.
Modeling Subcommittee, Technical Fisheries Committee. 2002. Summary status of Lake Trout and Lake Whitefish populations in the 1836 Treaty-Ceded waters of Lakes Superior, Huron and Michigan in 2000, with recommended yield and effort levels for 2001. Technical Fisheries Committee, 1836 Treaty-Ceded Waters of Lakes Superior, Huron, and Michigan. Available: http://www.michigan.gov/greatlakesconsentdecree. (May 2017).
Moll, R. A., C. Sellinger, E. S. Rutherford, J. L. Johnson, M. R. Fainter, and J. E. Gannon. 2013. The Great Lakes: an overview of their formation, geology, physics, and chemistry. Pages $3-30$ in W. W. Taylor, A. Lynch, and N. Leonard, editors. Great Lakes fisheries policy and management: a binational perspective, 2nd edition. Michigan State University Press, East Lansing.
Morrissey, M. B., C. D. Suski, K. R. Esseltine, and B. L. Tufts. 2005. Incidence and consequences of decompression in Smallmouth Bass after live-release angling tournaments. Transactions of the American Fisheries Society 134:1038-1047.
Muir, A. M., C. R. Bronte, M. S. Zimmerman, H. R. Quinlan, J. D. Glase, and C. C. Krueger. 2014. Ecomorphological diversity of Lake Trout at Isle Royale, Lake Superior. Transactions of the American Fisheries Society 143:972-987.
Muoneke, M. I., and W. M. Childress. 1994. Hooking mortality: a review for recreational fisheries. Reviews in Fisheries Science 2:123-156.
Ng, E. L., J. P. Fredericks, and M. C. Quist. 2015. Effects of gill-net trauma, barotrauma, and deep release on postrelease mortality of Lake Trout. Journal of Fish and Wildlife Management 6:265-277.
Paukert, C. P., J. A. Klammer, R. B. Pierce, and T. D. Simonson. 2001. An overview of Northern Pike regulations in North America. Fisheries 26(6):6-13.
Paukert, C. P., M. McInerny, and R. Schultz. 2007. Historical trends in creel limits, length-based limits, and season restrictions for black basses in the United States and Canada. Fisheries 32:62-72.
Persons, S. E., and S. A. Hirsch. 1994. Hooking mortality of Lake Trout angled through ice by jigging and set-lining. North American Journal of Fisheries Management 14:664-668.
Pollock, K. H., and W. E. PineIII . 2007. The design and analysis of field studies to estimate catch-and-release mortality. Fisheries Management and Ecology 14:123-130.
Quinn, T. J. II, and R. B. Deriso. 1999. Quantitative fish dynamics. Oxford University Press, New York.
R Core Team. 2016. A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna.
Schmalz, P. J., M. J. Hansen, M. E. Holey, P. C. McKee, and M. L. Toneys. 2002. Lake Trout movements in northwestern Lake Michigan. North American Journal of Fisheries Management 22:737-749.
Talmage, P. J., and D. F. Staples. 2011. Mortality of Walleyes angled from the deep waters of Rainy Lake, Minnesota. North American Journal of Fisheries Management 31:826-831.
Tracey, S. R., K. Hartmann, M. Leef, and J. McAllister. 2016. Captureinduced physiological stress and postrelease mortality for southern Bluefin Tuna (Thunnus maccoyii) from a recreational fishery. Canadian Journal of Fisheries and Aquatic Sciences 73:1547-1556.
Westerman, F. A. 1932. The deepwater trap net and its relation to Great Lakes fisheries. Transactions of the American Fisheries Society 62:64-71.
Wilde, G. R. 2009. Does venting promote survival of released fish? Fisheries 34:20-28.
Zar, J. H. 1999. Biostatistical analysis, 4th edition. Prentice Hall, Upper Saddle River, New Jersey.

## Appendix: Detailed Data

TABLE A.1. Marascuilo procedure for all pairwise comparisons of tag return rates according to angler handling times ( $<1.0,1.0-1.5,1.5-2.0,2.0-2.5$, and $>2.5$ min ) for Lake Trout tagged in Lake Superior. Marascuilo test statistics include absolute difference in proportions (a), critical value ( $r$ ), and statistical significance at $\alpha=0.05$ (yes or no).

| Comparison pair | $a$ | $r$ | $a-r$ | Significance |
| :--- | :---: | :---: | :---: | :---: |
| $<1.0$ min versus $1.0-1.5 \mathrm{~min}$ | 0.011 | 0.102 | No |  |
| $<1.0$ min versus $1.5-2.0 \mathrm{~min}$ | 0.013 | 0.117 | -0.091 | No |
| $<1.0$ min versus $2.0-2.5 \mathrm{~min}$ | 0.045 | 0.166 | -0.104 | No |
| $<1.0$ min versus $>2.5 \mathrm{~min}$ | 0.035 | 0.210 | -0.121 | No |
| $1.0-1.5$ min versus $1.5-2.0 \mathrm{~min}$ | 0.002 | 0.098 | No |  |
| $1.0-1.5$ min versus $2.0-2.5 \mathrm{~min}$ | 0.034 | 0.153 | -0.096 | No |
| $1.0-1.5$ min versus $>2.5 \mathrm{~min}$ | 0.046 | 0.098 | -0.119 | No |
| $1.5-2.0$ min versus $2.0-2.5 \mathrm{~min}$ | 0.032 | 0.163 | No |  |
| $2.0-2.5$ min versus $>2.5 \mathrm{~min}$ | 0.080 | 0.230 | -0.130 | No |

TABLE A.2. Marascuilo procedure for all pairwise comparisons of tag return rates according to angler handling times ( $<1.0,1.0-1.5,1.5-2.0,2.0-2.5$, and $>2.5$ $\mathrm{min})$ for Lake Trout tagged in Lake Huron. Marascuilo test statistics include absolute difference in proportions (a), critical value ( $r$ ), and statistical significance at $\alpha=0.05$ (yes or no).

| Comparison pair | $a$ | $r$ | $a-r$ | Significance |
| :--- | :---: | :---: | :---: | :---: |
| $<1.0$ min versus $1.0-1.5 \mathrm{~min}$ | 0.003 | 0.063 | -0.060 | No |
| $<1.0$ min versus $1.5-2.0 \mathrm{~min}$ | 0.016 | 0.097 | -0.081 | No |
| $<1.0$ min versus $2.0-2.5 \mathrm{~min}$ | 0.054 | 0.161 | -0.107 | No |
| $<1.0$ min versus $>2.5$ min | 0.038 | 0.152 | -0.113 | No |
| $1.0-1.5$ min versus $1.5-2.0 \mathrm{~min}$ | 0.019 | 0.098 | No |  |
| $1.0-1.5$ min versus $2.0-2.5 \mathrm{~min}$ | 0.057 | 0.162 | No |  |
| $1.0-1.5$ min versus $>2.5 \mathrm{~min}$ | 0.041 | 0.098 | No |  |
| $1.5-2.0$ min versus $2.0-2.5 \mathrm{~min}$ | 0.038 | 0.177 | -0.105 | No |
| $2.0-2.5$ min versus $>2.5 \mathrm{~min}$ | 0.015 | 0.212 | -0.140 | No |

TABLE A.3. Marascuilo procedure for all pairwise comparisons of tag return rates by angler fishing method for Lake Trout tagged in Lake Superior. Fishing methods are described in Table 1. Marascuilo test statistics include absolute difference in proportions ( $a$ ), critical value $(r)$, and statistical significance at $\alpha=0.05$ (yes or no).

| Comparison pair | $a$ | $r$ | $a-r$ | Significance |
| :--- | :---: | :---: | :---: | :---: |
| BOB versus DR-NR | 0.024 | 0.147 | -0.123 | No |
| BOB versus DR-REL | 0.015 | 0.066 | -0.050 | No |
| BOB versus WIRE | 0.059 | 0.089 | -0.030 | No |
| DR-NR versus DR-REL | 0.008 | 0.141 | -0.133 | No |
| DR-NR versus WIRE | 0.082 | 0.153 | No |  |
| DR-REL versus WIRE | 0.074 | 0.079 | -0.071 | No |

TABLE A.4. Marascuilo procedure for all pairwise comparisons of tag return rates by angler fishing method for Lake Trout tagged in Lake Huron. Fishing methods are described in Table 1. Marascuilo test statistics include absolute difference in proportions ( $a$ ), critical value ( $r$ ), and statistical significance at $\alpha=0.05$ (yes or no).

| Comparison pair | $a$ | $r$ | $a-r$ | Significance |
| :--- | :---: | :---: | :---: | :---: |
| SURF versus DR-REL | 0.030 | 0.084 | -0.054 | No |
| SURF versus WIRE/LC | 0.023 | 0.117 | -0.094 | No |
| DR-REL versus WIRE/LC | 0.053 | 0.085 | -0.032 | No |

TABLE A.5. Marascuilo procedure for all pairwise comparisons of tag return rates by play time interval $(\mathrm{p} 1=<1 \mathrm{~min} ; \mathrm{p} 2=1-2 \mathrm{~min} ; \mathrm{p} 3=2-3 \mathrm{~min} ; \mathrm{p} 4=3-4$ $\min ; p 5=4-5 \mathrm{~min} ; \mathrm{p} 6=>5 \mathrm{~min}$ ) for each fishing method for Lake Trout tagged in Lake Superior. Fishing methods are described in Table 1. Marascuilo test statistics include absolute difference in proportions (a), critical value ( $r$ ), and statistical significance (Sig) at $\alpha=0.05$ (yes or no). No fish were caught by the BOB method with play times greater than 4 min or by DR-NR and WIRE with play times less than 1 min .

| Comparison pair | BOB |  |  |  | DR-NR |  |  |  | DR-REL |  |  |  | WIRE |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $a$ | $r$ | $a-r$ | Sig | $a$ | $r$ | $a-r$ | Sig | $a$ | $r$ | $a-r$ | Sig | $a$ | $r$ | $a-r$ | Sig |
| p 1 versus p 2 | 0.19 | 0.41 | -0.22 | No |  |  |  |  | 0.30 | 0.16 | 0.14 | Yes |  |  |  |  |
| p1 versus p3 | 0.18 | 0.37 | -0.19 | No |  |  |  |  | 0.34 | 0.08 | 0.25 | Yes |  |  |  |  |
| p 1 versus p 4 | 0.31 | 0.56 | -0.25 | No |  |  |  |  | 0.32 | 0.08 | 0.24 | Yes |  |  |  |  |
| p 1 versus p 5 |  |  |  |  |  |  |  |  | 0.30 | 0.12 | 0.18 | Yes |  |  |  |  |
| p1 versus p6 |  |  |  |  |  |  |  |  | 0.30 | 0.11 | 0.19 | Yes |  |  |  |  |
| p 2 versus p 3 | 0.01 | 0.19 | -0.18 | No | 0.05 | 0.37 | -0.33 | No | 0.04 | 0.18 | -0.13 | No | 0.03 | 0.73 | -0.70 | No |
| p 2 versus p 4 | 0.12 | 0.46 | -0.34 | No | 0.09 | 0.43 | -0.34 | No | 0.02 | 0.17 | -0.15 | No | 0.03 | 0.70 | -0.67 | No |
| p2 versus p 5 |  |  |  |  | 0.20 | 0.64 | -0.44 | No | 0.01 | 0.20 | -0.19 | No | 0.03 | 0.69 | -0.66 | No |
| p2 versus p6 |  |  |  |  | 0.52 | 0.72 | -0.20 | No | 0.00 | 0.19 | -0.19 | No | 0.00 | 0.68 | -0.68 | No |
| p3 versus p4 | 0.13 | 0.42 | -0.30 | No | 0.13 | 0.42 | -0.28 | No | 0.02 | 0.11 | -0.09 | No | 0.07 | 0.34 | -0.27 | No |
| p3 versus p5 |  |  |  |  | 0.25 | 0.63 | -0.38 | No | 0.04 | 0.15 | -0.11 | No | 0.06 | 0.32 | -0.26 | No |
| p3 versus p6 |  |  |  |  | 0.57 | 0.71 | -0.15 | No | 0.04 | 0.14 | -0.10 | No | 0.04 | 0.30 | -0.26 | No |
| p4 versus p 5 |  |  |  |  | 0.11 | 0.66 | -0.55 | No | 0.01 | 0.14 | -0.13 | No | 0.00 | 0.25 | -0.24 | No |
| p4 versus p6 |  |  |  |  | 0.43 | 0.74 | -0.31 | No | 0.02 | 0.13 | -0.11 | No | 0.03 | 0.22 | -0.19 | No |
| p5 versus p6 |  |  |  |  | 0.32 | 0.88 | -0.56 | No | 0.01 | 0.16 | -0.16 | No | 0.03 | 0.19 | -0.17 | No |

TABLE A.6. Marascuilo procedure for all pairwise comparisons of tag return rates by play time interval ( $\mathrm{p} 1=<1 \mathrm{~min} ; \mathrm{p} 2=1-2 \mathrm{~min} ; \mathrm{p} 3=2-3 \mathrm{~min} ; \mathrm{p} 4=3-4$ min ; p5 $=4-5 \mathrm{~min}$; p6 $=>5 \mathrm{~min}$ ) for each fishing method for Lake Trout tagged in Lake Huron. Fishing methods are described in Table 1. Marascuilo test statistics include absolute difference in proportions (a), critical value ( $r$ ), and statistical significance (Sig) at $\alpha=0.05$ (yes or no).

| Pair | SURF |  |  |  | DR-REL |  |  |  | WIRE/LC |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $a$ | $r$ | $a-r$ | Sig | $a$ | $r$ | $a-r$ | Sig | $a$ | $r$ | $a-r$ | Sig |
| p 1 versus p 2 | 0.05 | 0.17 | -0.12 | No | 0.04 | 0.11 | -0.07 | No | 0.07 | 0.23 | -0.16 | No |
| p1 versus p3 | 0.17 | 0.36 | -0.19 | No | 0.03 | 0.12 | -0.09 | No | 0.06 | 0.19 | -0.13 | No |
| p1 versus p4 | 0.05 | 0.16 | -0.11 | No | 0.06 | 0.12 | -0.06 | No | 0.27 | 0.38 | -0.11 | No |
| p1 versus p 5 | 0.11 | 0.35 | -0.24 | No | 0.04 | 0.15 | -0.11 | No | 0.06 | 0.20 | -0.14 | No |
| p1 versus p6 | 0.00 | 0.00 | 0.00 | No | 0.07 | 0.13 | -0.06 | No | 0.06 | 0.20 | -0.14 | No |
| p2 versus p3 | 0.11 | 0.40 | -0.28 | No | 0.01 | 0.07 | -0.06 | No | 0.01 | 0.30 | -0.29 | No |
| p 2 versus p 4 | 0.01 | 0.23 | -0.23 | No | 0.02 | 0.06 | -0.04 | No | 0.20 | 0.44 | -0.25 | No |
| p2 versus p5 | 0.06 | 0.39 | -0.33 | No | 0.00 | 0.11 | -0.11 | No | 0.01 | 0.31 | -0.30 | No |
| p2 versus p6 | 0.05 | 0.17 | -0.12 | No | 0.02 | 0.08 | -0.06 | No | 0.01 | 0.31 | -0.30 | No |
| p3 versus p4 | 0.12 | 0.39 | -0.27 | No | 0.03 | 0.08 | -0.04 | No | 0.21 | 0.43 | -0.22 | No |
| p3 versus p5 | 0.06 | 0.50 | -0.44 | No | 0.01 | 0.12 | -0.11 | No | 0.00 | 0.28 | -0.27 | No |
| p3 versus p6 | 0.17 | 0.36 | -0.19 | No | 0.04 | 0.09 | -0.05 | No | 0.00 | 0.28 | -0.27 | No |
| p 4 versus p 5 | 0.06 | 0.38 | -0.32 | No | 0.02 | 0.12 | -0.10 | No | 0.20 | 0.43 | -0.23 | No |
| p4 versus p6 | 0.05 | 0.16 | -0.11 | No | 0.01 | 0.09 | -0.08 | No | 0.20 | 0.43 | -0.23 | No |
| p5 versus p6 | 0.11 | 0.35 | -0.24 | No | 0.02 | 0.13 | -0.10 | No | 0.00 | 0.29 | -0.29 | No |

TABLE A.7. Marascuilo procedure for all pairwise comparisons of tag return rates by capture depth interval (m) for all fishing methods combined for Lake Trout tagged in Lake Superior. Fishing methods are described in Table 1. Marascuilo test statistics include absolute difference in proportions (a), critical value ( $r$ ), and statistical significance at $\alpha=0.05$ (yes or no).

| Comparison pair | $a$ | $r$ | $a-r$ | Significance |
| :---: | :---: | :---: | :---: | :---: |
| $<20 \mathrm{~m}$ versus $20-30 \mathrm{~m}$ | 0.003 | 0.6 | -0.597 | No |
| $<20 \mathrm{~m}$ versus $30-40 \mathrm{~m}$ | 0.04 | 0.422 | -0.382 | No |
| $<20 \mathrm{~m}$ versus $40-50 \mathrm{~m}$ | 0.011 | 0.335 | -0.323 | No |
| $<20 \mathrm{~m}$ versus $50-60 \mathrm{~m}$ | 0.084 | 0.332 | -0.248 | No |
| $<20 \mathrm{~m}$ versus $60-70 \mathrm{~m}$ | 0.055 | 0.344 | -0.288 | No |
| $<20 \mathrm{~m}$ versus 70-80 m | 0.097 | 0.348 | -0.251 | No |
| $<20 \mathrm{~m}$ versus $>80 \mathrm{~m}$ | 0.049 | 0.339 | -0.29 | No |
| $20-30 \mathrm{~m}$ versus $30-40 \mathrm{~m}$ | 0.037 | 0.57 | -0.533 | No |
| $20-30 \mathrm{~m}$ versus $40-50 \mathrm{~m}$ | 0.008 | 0.509 | -0.501 | No |
| $20-30 \mathrm{~m}$ versus $50-60 \mathrm{~m}$ | 0.081 | 0.507 | -0.427 | No |
| $20-30 \mathrm{~m}$ versus $60-70 \mathrm{~m}$ | 0.052 | 0.515 | -0.463 | No |
| $20-30 \mathrm{~m}$ versus $70-80 \mathrm{~m}$ | 0.093 | 0.518 | -0.425 | No |
| $20-30 \mathrm{~m}$ versus $>80 \mathrm{~m}$ | 0.045 | 0.512 | -0.467 | No |
| $30-40 \mathrm{~m}$ versus $40-50 \mathrm{~m}$ | 0.029 | 0.278 | -0.249 | No |
| $30-40 \mathrm{~m}$ versus $50-60 \mathrm{~m}$ | 0.044 | 0.274 | -0.231 | No |
| $30-40 \mathrm{~m}$ versus $60-70 \mathrm{~m}$ | 0.015 | 0.288 | -0.273 | No |
| $30-40 \mathrm{~m}$ versus $70-80 \mathrm{~m}$ | 0.057 | 0.294 | -0.237 | No |
| $30-40 \mathrm{~m}$ versus $>80 \mathrm{~m}$ | 0.008 | 0.283 | -0.274 | No |
| $40-50 \mathrm{~m}$ versus $50-60 \mathrm{~m}$ | 0.073 | 0.096 | -0.024 | No |
| $40-50 \mathrm{~m}$ versus $60-70 \mathrm{~m}$ | 0.044 | 0.131 | -0.087 | No |
| $40-50 \mathrm{~m}$ versus $70-80 \mathrm{~m}$ | 0.085 | 0.142 | -0.056 | No |
| $40-50 \mathrm{~m}$ versus $>80 \mathrm{~m}$ | 0.037 | 0.118 | -0.08 | No |
| $50-60 \mathrm{~m}$ versus $60-70 \mathrm{~m}$ | 0.029 | 0.124 | -0.095 | No |
| $50-60 \mathrm{~m}$ versus $70-80 \mathrm{~m}$ | 0.013 | 0.135 | -0.123 | No |
| $50-60 \mathrm{~m}$ versus $>80 \mathrm{~m}$ | 0.035 | 0.11 | -0.074 | No |
| $60-70 \mathrm{~m}$ versus $70-80 \mathrm{~m}$ | 0.042 | 0.162 | -0.12 | No |
| $60-70 \mathrm{~m}$ versus $>80 \mathrm{~m}$ | 0.007 | 0.141 | -0.135 | No |
| $70-80 \mathrm{~m}$ versus $>80 \mathrm{~m}$ | 0.048 | 0.151 | -0.103 | No |

TABLE A.8. Marascuilo procedure for all pairwise comparisons of tag return rates by capture depth interval (m) for all fishing methods combined for Lake Trout tagged in Lake Huron. Fishing methods are described in Table 1. Marascuilo test statistics include absolute difference in proportions (a), critical value ( $r$ ), and statistical significance at $\alpha=0.05$ (yes or no).

| Comparison pair | $a$ | $r$ | $a-r$ | Significance |
| :--- | :---: | :---: | :---: | :---: |
| $<10 \mathrm{~m}$ versus $10-20 \mathrm{~m}$ | 0.031 | 0.186 | No |  |
| $<10 \mathrm{~m}$ versus $20-30 \mathrm{~m}$ | 0.037 | 0.171 | -0.155 | No |
| $<10 \mathrm{~m}$ versus $30-40 \mathrm{~m}$ | 0.028 | 0.173 | -0.134 | No |
| $<10 \mathrm{~m}$ versus $40-50 \mathrm{~m}$ | 0.024 | 0.183 | No |  |
| $<10 \mathrm{~m}$ versus $>50 \mathrm{~m}$ | 0.037 | 0.387 | -0.145 | No |
| $10-20 \mathrm{~m}$ versus $20-30 \mathrm{~m}$ | 0.068 | 0.086 | -0.35 | No |
| $10-20 \mathrm{~m}$ versus $30-40 \mathrm{~m}$ | 0.059 | 0.091 | No |  |
| $10-20 \mathrm{~m}$ versus $40-50 \mathrm{~m}$ | 0.055 | 0.108 | No |  |
| $10-20 \mathrm{~m}$ versus $>50 \mathrm{~m}$ | 0.006 | 0.358 | No |  |
| $20-30 \mathrm{~m}$ versus $30-40 \mathrm{~m}$ | 0.009 | 0.054 | -0.032 | No |
| $20-30 \mathrm{~m}$ versus $40-50 \mathrm{~m}$ | 0.013 | 0.079 | No |  |
| $20-30 \mathrm{~m}$ versus $>50 \mathrm{~m}$ | 0.074 | 0.35 | No |  |
| $30-40 \mathrm{~m}$ versus $40-50 \mathrm{~m}$ | 0.004 | 0.084 | -0.351 | No |
| $30-40 \mathrm{~m}$ versus $>50 \mathrm{~m}$ | 0.065 | 0.351 | -0.067 | No |
| $40-50 \mathrm{~m}$ versus $>50 \mathrm{~m}$ | 0.061 | 0.356 | -0.276 | No |


[^0]:    *Corresponding author: sitars@michigan.gov
    ${ }^{1}$ Retired.
    Received January 26, 2017; accepted April 29, 2017

